

MEMORANDUM



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CC:

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LAKE WELLS: NDVI, NDWI AND ET CALCULATIONS

Hydrobiology has completed the GIS tasks and extraction of summary data for Normalised Difference Vegetation Index (NDVI) and Normalised Difference Wetness Index (NDWI or Wetness Index) at Lake Wells, WA, the location of Australian Potash Limited's proposed Lake Wells potash project. The purpose of this analysis was to provide input into an assessment of possible impacts on vegetation from elements of the proposed Lake Wells project. Specifically, the analysis of spectral data is proposed as one source of information to help evaluate the possible presence of groundwater dependent vegetation in the project locality.

The years 2006 and 2008 were chosen for analysis of Landsat 4-5 imagery (NDVI and NDWI) based on suitable dry season conditions and availability of high-quality cloud-free imagery. NDVI provides a reliable measure of chlorophyll content or greenness of the vegetation. It is suggested that low seasonal variance in NDVI values is typically a signature of vegetation that has access to groundwater, and this relationship can often be more apparent at the end of the dry season when water is limited (e.g. Barron et al 2012).

An assessment was also conducted of Evapotranspiration (ET) for the same years (2006 and 2008) using the CSIRO MODIS reflectance-based scaling evapotranspiration (CMRSET) data set. Groundwater-dependent vegetation (GDE) is commonly associated with higher rates of ET, hence by calculating ET it may be possible to identify potential GDEs, especially when taken in concert with the NDVI and NDWI measures (e.g. Guerschman et al. 2009).

The use of remote sensing to assess vegetation function has recently become an established technique. "Remote sensing provides a robust and spatially explicit means to assess not only vegetation structure and function but also relationships amongst these and climate variables" (Eamus et al. 2015).

Brief Methods

NDVI and NDWI Method

The general approach to identification of potential GDEs followed Barron et al. (2012) – "Mapping groundwater-dependent ecosystems using remote sensing measures of vegetation and moisture dynamics". This involved using multi-spectral imagery to derive NDVI and NDWI parameters using the red, near infrared and short-wave infrared bands (as described in Barron et al. 2012). Landsat imagery at a spatial resolution of 30 m × 30 m was obtained from the USGS Earth Explorer web service for the Lake Wells study area ([Figure 1](#)). Images from the end of the wet season (Feb-April) until the end of the dry season (Sept-Nov) were obtained for two years (2006 and 2008). These years were chosen based on a combination of an extended dry spell of several months and suitable cloud-free imagery available for the whole study area. Rainfall data were sourced from the



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nearest Bureau of Meteorology meteorological station: the Laverton met station is located approximately 180 km southwest of Lake Wells. Imagery was obtained for a temporal frequency of monthly where possible.

Landsat imagery from the Landsat 4-5 archive was determined to be the most appropriate for this analysis as it contained a suitable temporal and spatial coverage. The Landsat 7 imagery contains a sensor error that may have made comparison between the older Landsat 4-5 and Landsat 7 imagery (which covers years 2009 to 2013) subject to error. The current Landsat 8 imagery is only processed for the study area for 2016 to present, which represents an unusually wet period.

Raw imagery in GeoTIFF format was downloaded from the USGS website and Bands 3, 4 and 5 ([Table 1](#)~~Table 4~~) were imported into the QGIS software package for processing. Each image was clipped to a standard coverage area ([Figure 1](#)~~Figure 4~~) and the NDVI and NDWI values calculated using Python scripting within the QGIS software. A vegetation community (Floristic community) map provided by Australian Potash Limited (APL) (produced by Botanica) was used to select zones for the generation of statistics by vegetation type for 2006 and 2008 ([Table 2](#)~~Table 2~~). [Figure 1](#)~~Figure 1~~ and [Table 4](#)~~Table 4~~ provide background information on the vegetation communities analysed. Descriptive statistics were generated for each vegetation type using Zonal Statistics module within QGIS.

An additional vegetation/land cover unit classification was created to distinguish the wetting lake bed areas from the more purely terrestrial vegetation units. The reason for this extra classification was to identify and qualify the lake bed signature so that this could be compared with the vegetation units in terms of ET, NDVI and NDWI. This was undertaken to determine if the lake bed was influencing or skewing the results. It can be difficult to delineate the lake bed from the surrounding area because it is typically covered, at least partially, with CD-CSSSF1 vegetation. Delineation of areas that were within the lake bed was achieved using the MODIS product - Water Observations from Space (WOfS), which maps the presence of surface water across Australia. The study area was mapped selecting for areas that have water at least 1% of the time (i.e. ~3 days a year). The resulting lake bed area was named 'Lake Bed CD-CSSSF1'.

Table 1 Landsat 4-5 Thematic mapper band information

Bands	Wavelength (micrometers)	Resolution (meters)
Band 3 - Red	0.63-0.69	30
Band 4 - Near Infrared (NIR)	0.76-0.90	30
Band 5 - Shortwave Infrared (SWIR) 1	1.55-1.75	30

Table 2 Number of images processed by year

Year	Number of NDVI images	Number of NDWI images	Total
2006	8	8	16
2008	7	7	14
TOTAL	15	15	30

Evapotranspiration Method

Estimates of actual evapotranspiration (AET) were calculated for the study area using satellite imagery from the 'CSIRO MODIS reflectance based scaling evapotranspiration' (CMRSET) data set (250 m resolution). This data set was developed by Guerschman et al. (2009) and it provides an estimate of AET across Australia, based on MODIS reflectance and short wave infra-red data, and gridded meteorological surfaces.

In brief, the CMRSET algorithm uses reflectance data from the MODIS satellite to calculate ET across the Australian continent. AET is calculated from potential ET (PET) by applying a 'crop factor' which incorporates the enhanced vegetation index (EVI) and global vegetation moisture index (GVMI). The algorithm was calibrated by comparing estimated AET with measured AET from seven eddy covariance towers around Australia covering a variety of landscapes (forest, savannah, grassland, floodplain and lake). CMRSET was further validated by comparing estimated AET with 'surrogate AET' (precipitation minus streamflow) in 227 unimpaired catchments around Australia Guerschman et al. (2009).

A cautious approach is required when attempting to make inferences about the presence of GDE from AET for several reasons. The first being that the amount of ET for a given vegetation type can be influenced by other factors such as vegetation health, leaf area index and how water tolerant the vegetation type is (Gonzalez 2015, Woods et al. 2016). Secondly the calibration method used for the CMRSET was conducted in areas with rainfall of greater than 250 mm and not in low rainfall areas like the study area. Thirdly, van Dijk et al. 2015 found that this method has a tendency to overestimate ET from salt lakes, however van Dijk et al showed that CMRSET method can provide reliable estimates ET in areas other than salt lakes.

Raw imagery in .nc format was downloaded from the NCI (National Computational Infrastructure) website and imported into QGIS software package for processing. A vegetation community (Floristic community) map provided by APL (produced by Botanica) was used to generate statistics by vegetation type for 2006 and 2008 (Table 4Table 4). Descriptive statistics were generated for each vegetation type using Zonal Statistics module within QGIS.

Table 3 Number of images processed by year

Year	Number of ET images
2006	9
2008	11
TOTAL	20

Calculation of Groundwater Evapotranspiration

Groundwater evapotranspiration (ETg) refers to the water losses from groundwater due to transpiration, direct water uptake through roots from GDEs, and direct evaporation (e.g. from any wet surface including soil or land surface). Groundwater-dependent vegetation is commonly associated with a comparatively higher rate of evapotranspiration (ETg), hence by identifying areas where ETg exceeds rainfall on an annual basis it is possible to predict potential GDEs (O'Grady et al. 2011). It is important to know that this method is a simplification of the system and does not include a direct measure of evaporation and assumes that 100% of the ETg comes from transpiration. Eamus et al. 2015 estimated that the average error associated with this method was about 12%, however it is likely to be much greater in environments where groundwater is expressed at the surface and or moist soil i.e. salt lakes and wetlands. In these types of environments there will be greater groundwater expression and hence higher ET, and it is highly likely that these high ET are not due to the presence of GDE but due to limitations of the method. Hence caution needs to be applied when making inference about GDEs associated with groundwater expressed at the surface. The rainfall data used for this calculation came from the Laverton rain gauge station which is 180 km from the study site, this data was used because it is the closest and most complete data set that was available. The distance of the rainfall data from the study site is another limitation of this method.

The spatial resolution of the ET data allows for a pixel size of 250m². The vegetation in this area can be highly patchy and may not completely fill a pixel, hence other components may be incorporated into the calculations (e.g. bare salt lake surface, open ground). This limitation needs to be considered when interpreting the ET results.

Groundwater evapotranspiration (ET_g) can be calculated from satellite imagery using NDVI and rainfall using the following formula in which NDVI* is the peak season normalised NDVI. It is important to note that these ET_g figures are estimates and to obtain more accurate results it is suggested that the model is calibrated using sites with in-situ measured ET.

$$E = (E_{max} - A_{min}) \times NDVI^*$$

NDVI* was calculated by subtracting the NDVI for the area that had no/the lowest amount of vegetation (NDVI_z, i.e. Lake Bed CD-CSSSF1') from the summer peak season NDVI for each vegetation unit (NDVI_p) and dividing this by NDVI at saturation (NDVI_m; the maximum value obtained by any vegetation unit), minus the NDVI_z, (Eamus et al. 2015).

$$NDVI^* = \frac{NDVI_p - NDVI_z}{NDVI_m - NDVI_z}$$

The following figures have been provided as a graphical presentation of the results:

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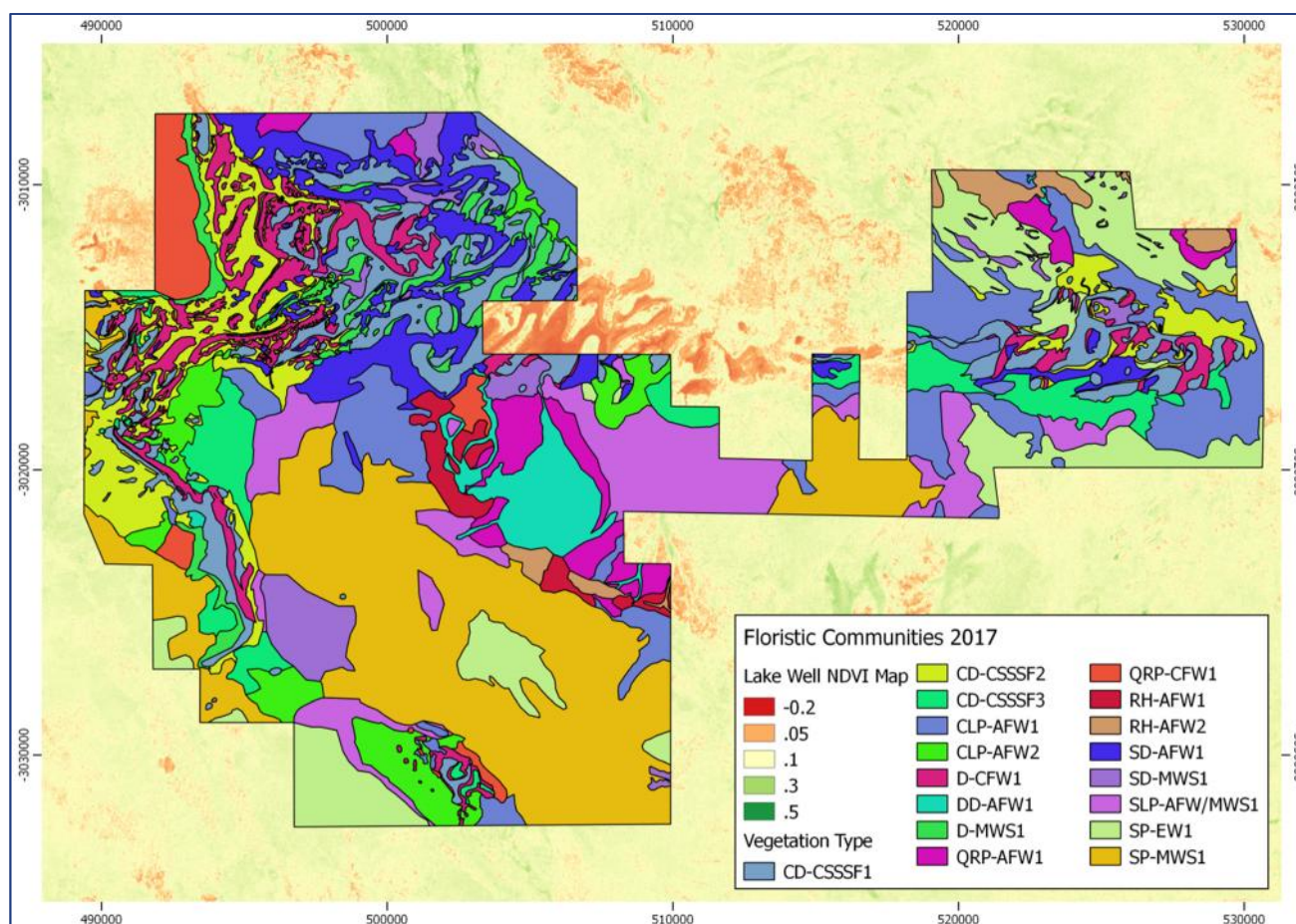


Figure 1 Map of vegetation communities used in the NDVI, NDWI and ET analysis

Table 4 Vegetation community descriptions

Landform	NVIS Vegetation Group	Vegetation Type	Vegetation Code
Closed Depression	Chenopod Shrublands, Samphire Shrublands and Forblands (MVG 22)	Low samphire shrubland of <i>Tecticornia indica</i> subsp. <i>bidens</i> / <i>Tecticornia</i> sp. Dennys Crossing (K.A. Shepherd & J. English KS522) in playa	CD-CSSSF1
		Mid heathland of <i>Cratystylis subspinescens</i> over low open chenopod shrubland of <i>Atriplex vesicaria</i> and open forbland of <i>Frankenia</i> spp. on playa edge	CD-CSSSF2
		Mid open shrubland of <i>Eremophila paisleyi</i> / <i>Lawrencia squamata</i> / <i>Lycium australis</i> over low open chenopod shrubland of <i>Atriplex</i> spp. and open forbland of <i>Frankenia</i> spp. on playa edge	CD-CSSSF3
Clay-Loam Plain	Acacia Forests and Woodlands (MVG 6)	Low open forest of <i>Acacia incurvaneura</i> over mid shrubland of <i>Eremophila margarethae</i> and low open tussock grassland of <i>Eriachne mucronata</i> / <i>Eragrostis eriopoda</i> on clay loam plain	CLP-AFW1
	Acacia Forests and Woodlands (MVG 6)	Low woodland of <i>Acacia caesaneura</i> over mid open shrubland of <i>A. burkittii</i> and mid chenopod shrubland of <i>Maireana pyramidata</i> / low open hummock grassland of <i>Triodia desertorum</i> on clay loam plain	CLP-AFW2
Drainage Depression	Acacia Forests and Woodlands (MVG 6)	Low open forest of <i>Acacia caesaneura</i> over mid open shrubland of <i>Senna artemisioides</i> subsp. <i>filifolia</i> and low open tussock grassland of <i>Eragrostis eriopoda</i> in drainage depression	DD-AFW1
Gypsum Dunefield	Casuarina Forests and Woodlands (MVG 8)	Low open forest of <i>Casuarina pauper</i> over tall open shrubland of <i>Acacia burkittii</i> and low sparse chenopod shrubland of <i>Atriplex vesicaria</i> on gypsum dune	D-CFW1
	Mallee Woodlands and Shrublands MVG 14)	Mid open mallee forest of <i>Eucalyptus gypsophila</i> over mid open shrubland of <i>Senna artemisioides</i> / <i>Eremophila</i> spp. and low open chenopod shrubland of <i>Atriplex vesicaria</i> on gypsum dune	D-MWS1
Quartz/ Rocky Plain	Acacia Forests and Woodlands (MVG 6)	Low open woodland of <i>Acacia caesaneura</i> / <i>A. incurvaneura</i> over mid open shrubland of <i>A. burkittii</i> / <i>Eremophila fraseri</i> and low open shrubland of <i>Ptilotus obovatus</i> / sparse tussock grassland of <i>Eragrostis eriopoda</i> on quartz/rocky plain	QRP-AFW1
	Casuarina Forests and Woodlands (MVG 8)	Low woodland of <i>Casuarina pauper</i> over mid shrubland of <i>Eremophila paisleyi</i> subsp. <i>paisleyi</i> / <i>Senna artemisioides</i> subsp. <i>filifolia</i> and low open shrubland of <i>Ptilotus obovatus</i> on quartz/rocky plain	QRP-CFW1
Rocky Hillslope	Acacia Forests and Woodlands	Low open forest of <i>Acacia quadrimarginea</i> over mid open shrubland of <i>Senna artemisioides</i> subsp. <i>filifolia</i> / <i>Senna</i> sp. <i>Meekatharra</i> (E. Bailey 1-26) and low open shrubland of <i>Ptilotus obovatus</i> on rocky hillslope	RH-AFW1

Landform	NVIS Vegetation Group	Vegetation Type	Vegetation Code
	(MVG 6)	Low woodland of <i>Acacia incurvaneura</i> over mid open shrubland of <i>Eremophila jucunda</i> and tussock grassland of <i>Eragrostis eriopoda</i> / <i>Eriachne mucronata</i> on rocky hillslope	RH-AFW2
Sand Dunefield	Acacia Forests and Woodlands (MVG 6)	Low woodland of <i>Acacia caesaneura</i> / <i>A. incurvaneura</i> over tall open shrubland of <i>Eremophila</i> spp./ <i>Senna</i> spp./ <i>Melaleuca interoris</i> and low open hummock grassland of <i>Triodia basedowii</i> / low open tussock grassland of <i>Eragrostis eriopoda</i> in dunefield	SD-AFW1
	Mallee Woodlands and Shrublands (MVG 14)	Mid mallee woodland of <i>Eucalyptus concinna</i> over low open shrubland of <i>Aluta maisonneuvei</i> subsp. <i>auriculata</i> / <i>Dodonaea viscosa</i> and low closed hummock grassland of <i>Triodia desertorum</i> in dunefield	SD-MWS1
Sand-Loam Plain	Acacia Forests and Woodlands/ Mallee Woodlands and Shrublands (MVG 6/ 14)	Low open forest of <i>Acacia caesaneura</i> / mid mallee woodland of <i>Eucalyptus lucasii</i> over mid open shrubland of <i>Eremophila latrobei</i> subsp. <i>glabra</i> and low hummock grassland of <i>Triodia desertorum</i> on sand-loam plain	SLP-AFW/MWS1
Sandplain	Eucalypt Woodlands (MVG 5)	Low woodland of <i>Eucalyptus gongylocarpa</i> over mid open shrubland of <i>Eremophila platythamnos</i> subsp. <i>exotrachys</i> and low hummock grassland of <i>Triodia desertorum</i> on sandplain	SP-EW1
	Mallee Woodlands and Shrublands (MVG 14)	Mid mallee shrubland of <i>Eucalyptus</i> spp. over mid open shrubland of <i>Acacia</i> spp. and low closed hummock grassland of <i>Triodia basedowii</i> on sandplain	SP-MWS1

While the present study has used methods consistent with Barron et al. (2012), it should be noted that the NDVI values returned between the two studies will differ due to vastly different vegetation communities and seasonal conditions. [Figure 2](#) provides an example of NDVI maps for Lake Wells and the Barron et al. (2012) study areas for May 2006, showing the “greener” conditions for the Swan Coastal Plain area when compared to dry marginal landscape of Lake Wells.

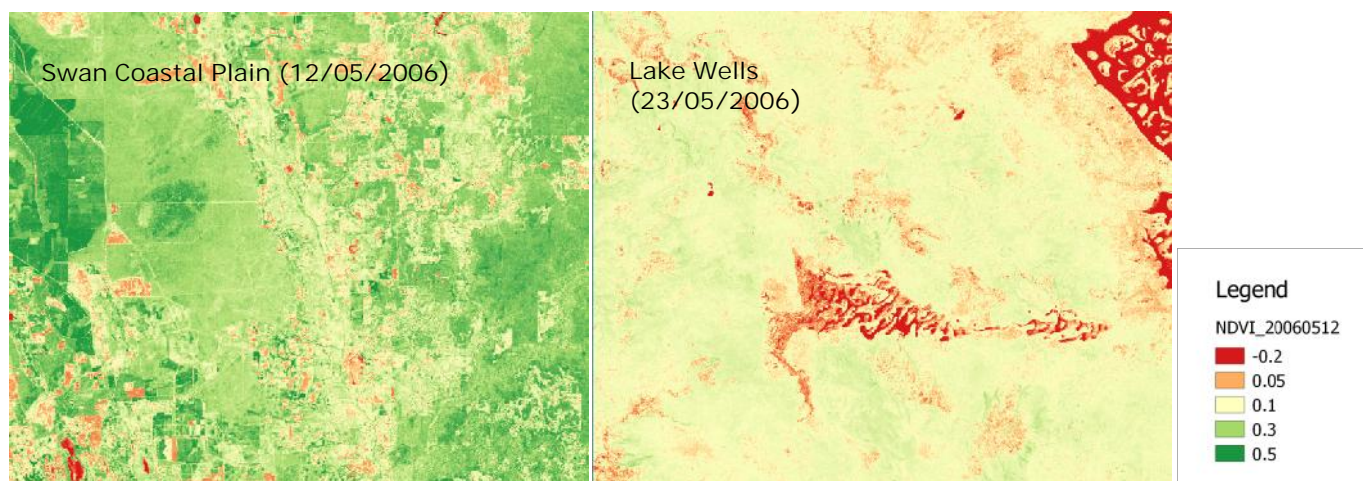


Figure 2 Comparison of NDVI values for Lake Wells (right hand side) and Swan Coastal Plain (left hand side) for 12/05/2006 and 23/05/2006 respectively

Results

NDVI and NDWI

Barron et al. (2012) assessed the presence of potential GDEs by plotting changes in NDVI values over the dry season, hypothesizing that those vegetation communities with the least change in greenness are most likely to be supplemented by water sources other than rainfall (i.e. groundwater or perched surface water). Following this method, [Figure 3](#) provides a plot of the late wet season NDVI value (x-axis) against the late dry season NDVI (y-axis) for each floristic community identified in the study area for each year (2006 and 2008). Vegetation units that deviate most from the 1:1 line are classified as 'fast-drying vegetation' (Barron et al, 2012) and are very unlikely to be groundwater dependent. Vegetation units with relatively high and unvarying NDVI values, which closely follow the 1:1 plot line are inferred to have a continuing source of water (i.e. are considered to be more likely to be groundwater dependent). Units with consistently low and unvarying NDVI may represent permanent water or wetland surfaces (if they also show high and unvarying NDWI signatures and high ET) or may correspond to sparse vegetation or bare soil (if they have lower NDWI and low cumulative ET). By ranking the ratio of NDVI values from late wet to end of dry season based on their distance to the 1:1 line we can identify which vegetation unit has the least variable NDVI value (see [Figure 4](#)).

This method has identified CD-CSSSF1 (Low samphire shrubland of *Tecticornia indica* subsp. *bidens*/ *Tecticornia* sp. Denny's Crossing (K.A. Shepherd & J. English KS522) in playa) and D-CFW1 (Low open *Casuarina* forest over *Acacia* shrubland and sparse chenopod shrubland) as having the least variable NDVI values across the dry season for both 2006 and 2008. There is a distinct possibility that the sparseness of vegetation, particularly in the CD-CSSSF1 community, is lowering the NDVI response over the dry season. The NDVI pixels are an average of 30 m × 30 m, which includes any bare ground between plants. Sparse vegetation would comprise a greater degree of non-variable substrate in the form of bare ground and/or dead litter material. Therefore the Barron et al. (2012) method is likely to require greater botanic interpretation when applied to the Lake Wells study area.

The relative change in NDVI values across the dry season (as measured by the distance from the 1:1 line) for the majority of vegetation units were greater in 2006 in comparison to 2008. For example, for D-CFW1 the distance unit was 0.05 in 2006 and 0.02 in 2008 (see [Figure 7](#) and [Figure 9](#)). A comparison of cumulative rainfall for both years ([Figure 5](#)) shows that 2008 had significantly lower rainfall throughout the year which may explain the reduced start of season NDVI values in the vegetation units. It is interesting to note that the only unit that didn't change between the two years was the Lake Bed CDSSS1, with NDVI values of -0.09 for both years. This indicates that NDVI values for Lake Bed CDSSS1 is not impacted by inter-annual rainfall (including associated groundwater recharge) and supports the theory that this area is displaying a predominantly lake bed signature. This relatively constant lake bed signature across seasons has been previously observed by Hydrobiology in similar NDVI assessment projects in inland arid Australia.

There was an early rainfall event in late 2008 on the 23/11/2008 which brought 40mm of rain and whilst the vast majority of the vegetation units responded to this rainfall event with an increase of NDVI, CD-CSSSF1, D-CFW1 and Lake Bed CDSSS1 all experienced declines in NDVI values. This can potentially be explained by the presence of pooling water in and around the lake bed which absorbs infrared light causing a reduction in NDVI, while in reality the 'greenness' of the vegetation may have actually increased or remained the same.

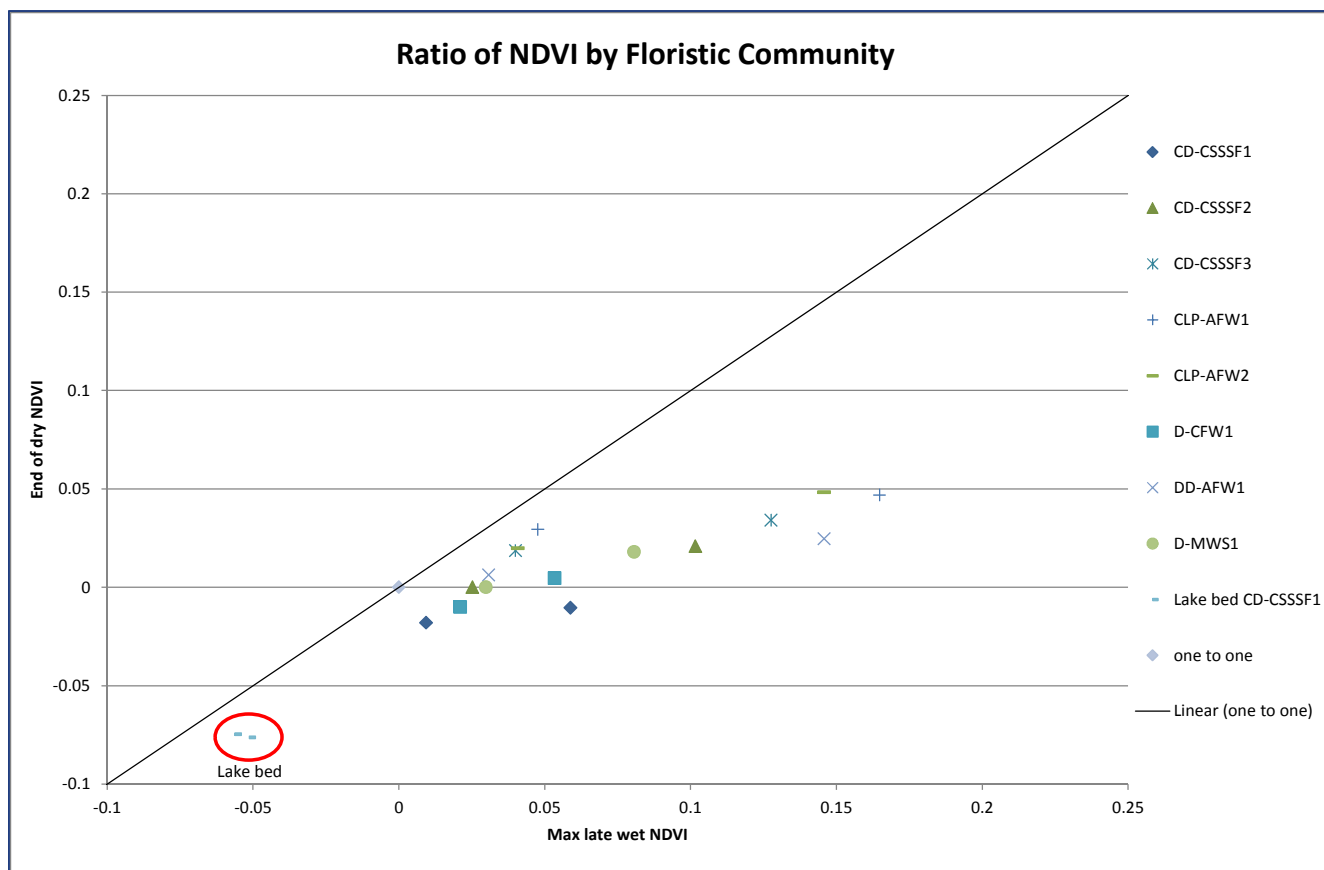


Figure 3 Ratio of NDVI values from late wet to end of dry season (2006 and 2008)

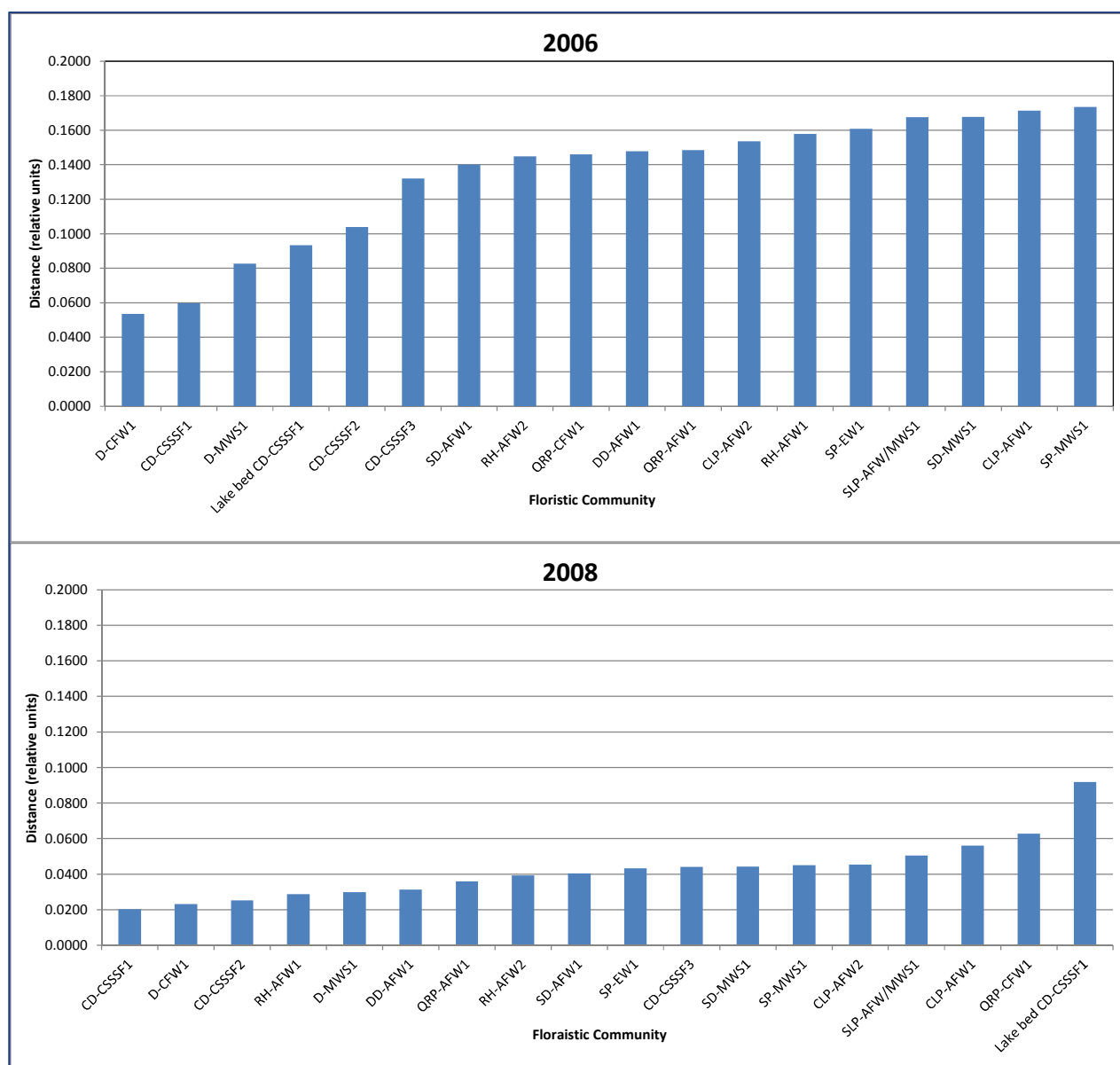


Figure 4 Ranking of 'distance' values from 1:1 (no change over dry season) for all floristic communities for 2006 and 2008

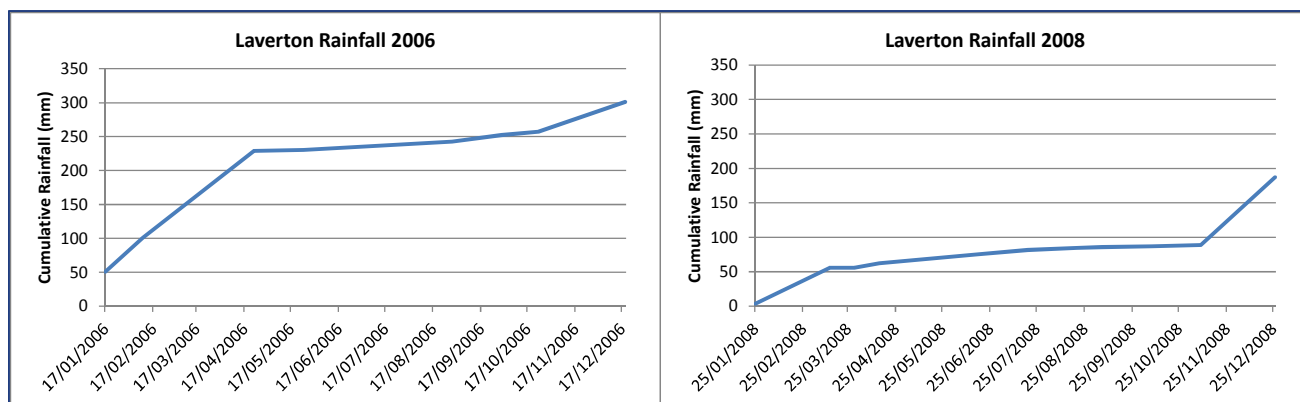


Figure 5 Cumulative rainfall for 2006 and 2008 (Laverton rain gauge station)

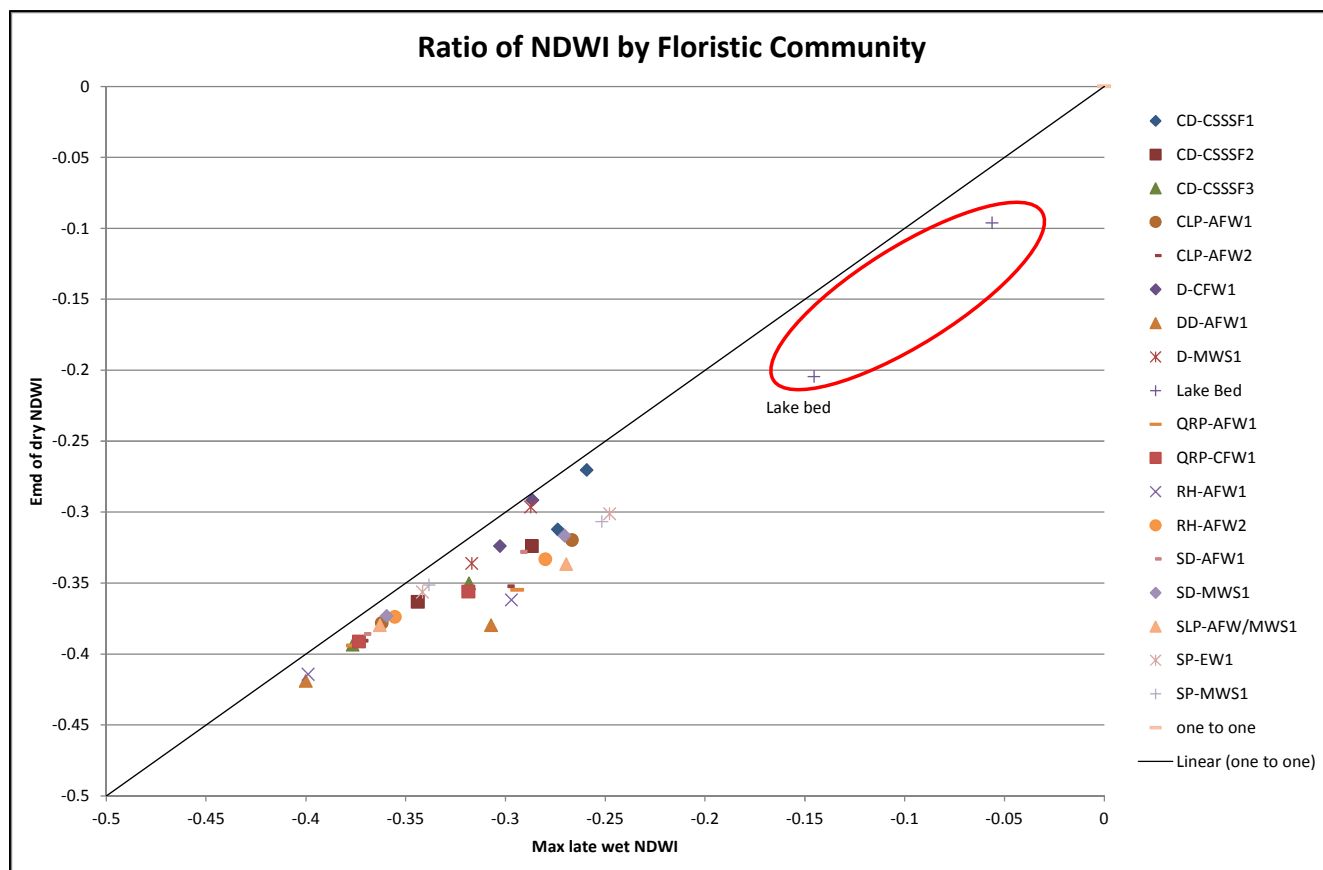


Figure 6 Ratio of NDWI values from late wet to end of dry seasons for 2006 and 2008

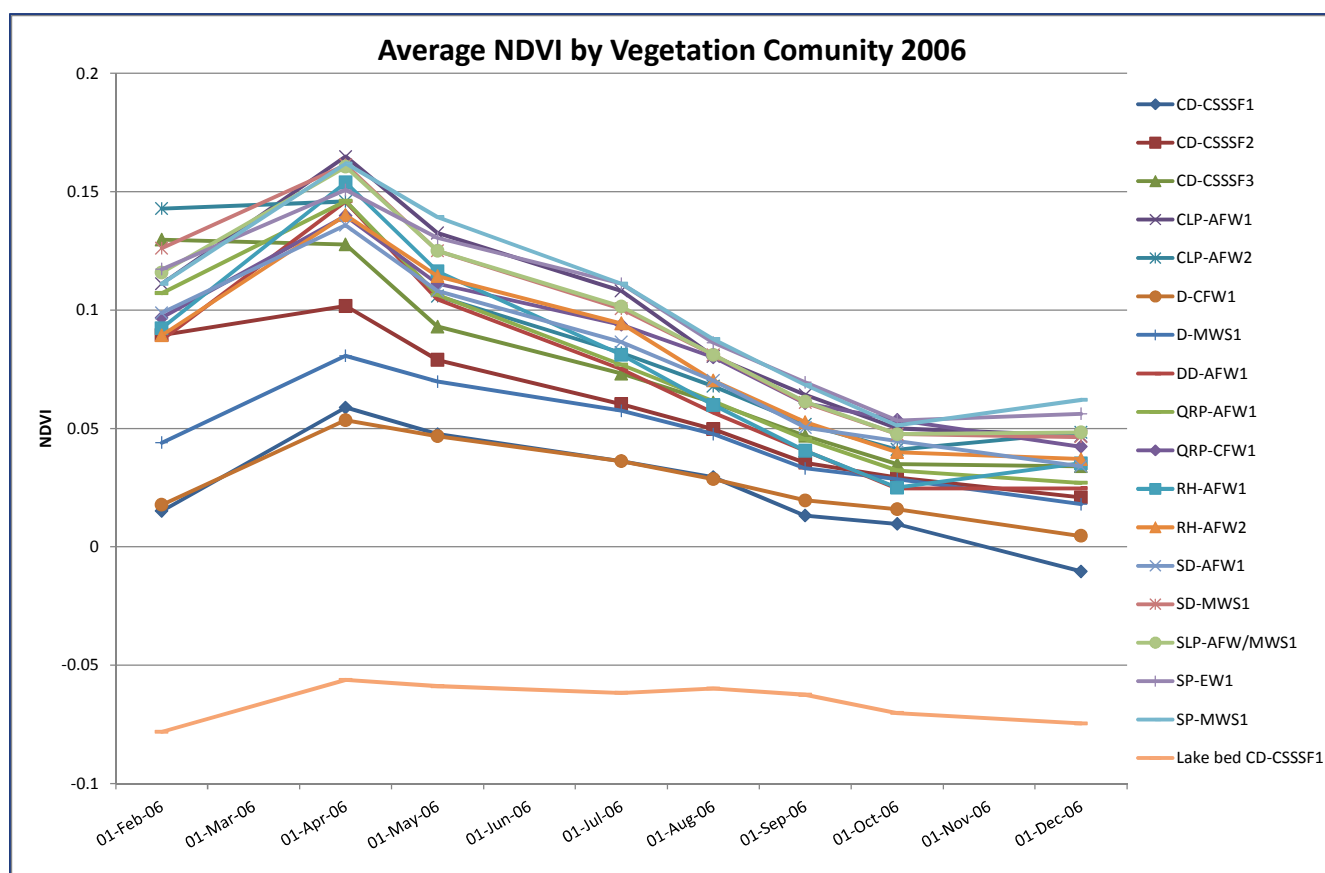


Figure 7 Average NDVI of each vegetation type for 2006

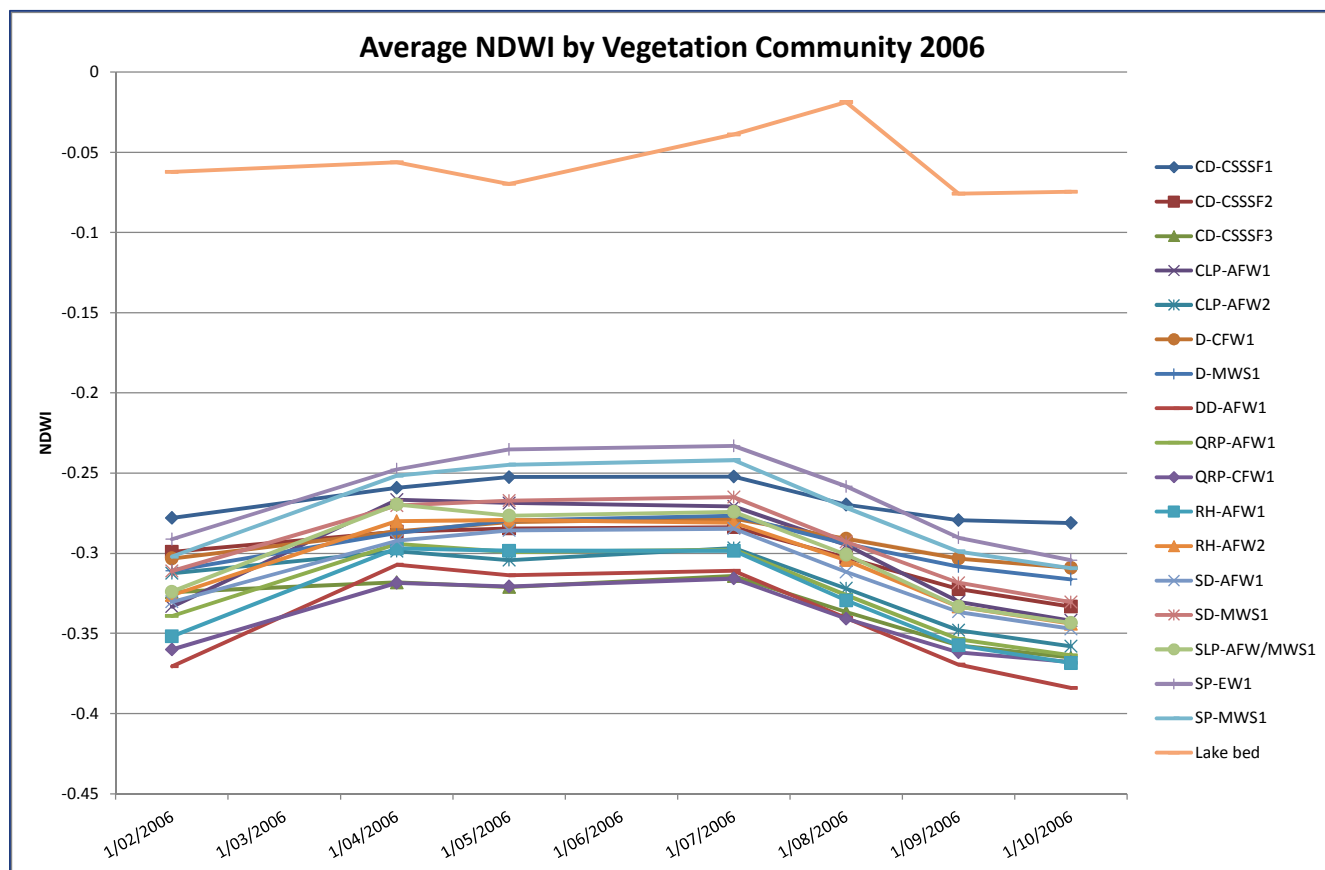


Figure 8 Average NDWI of each vegetation type for 2006

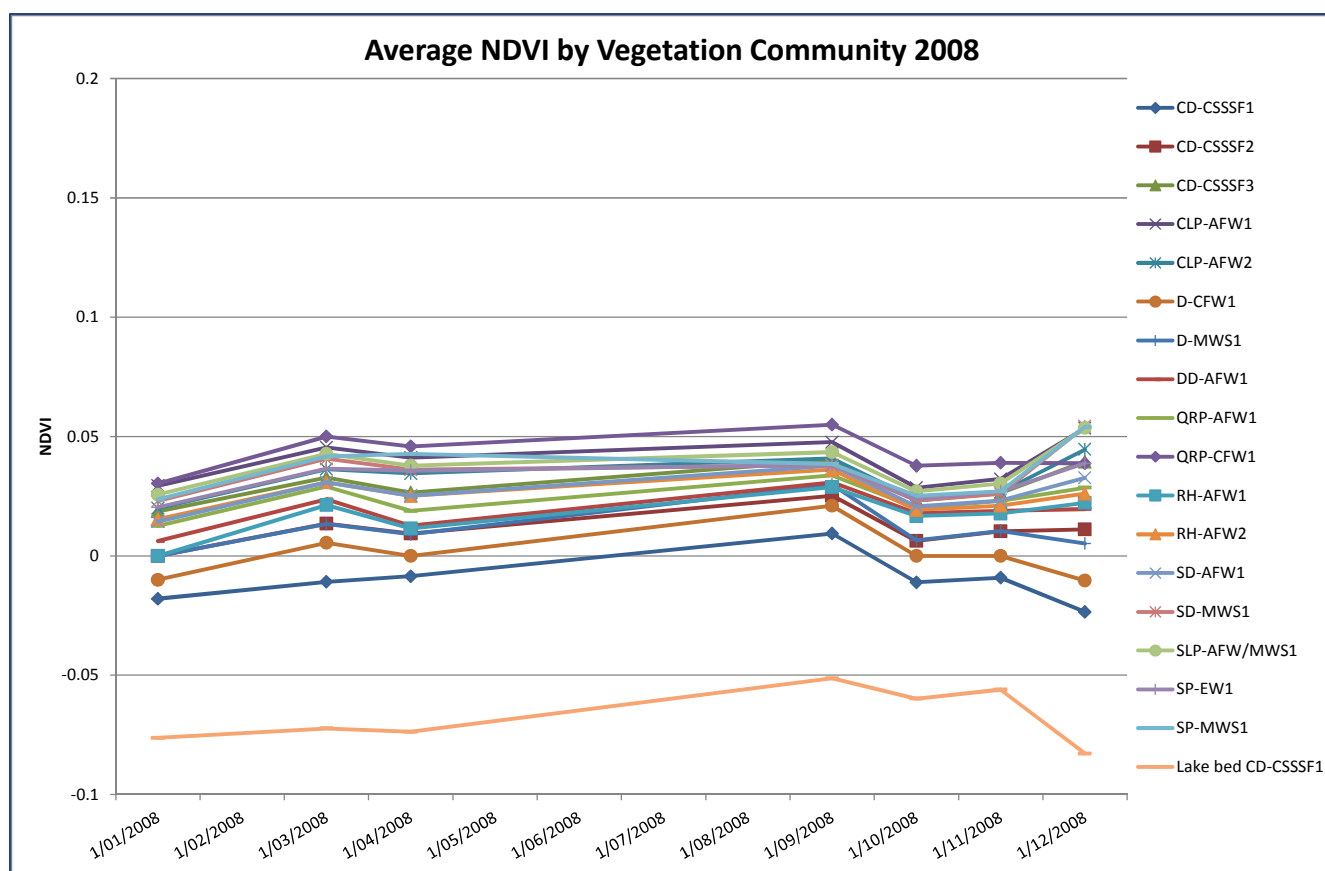


Figure 9 Average NDVI of each vegetation type for 2008 along with cumulative rainfall at Laverton rain gauge station

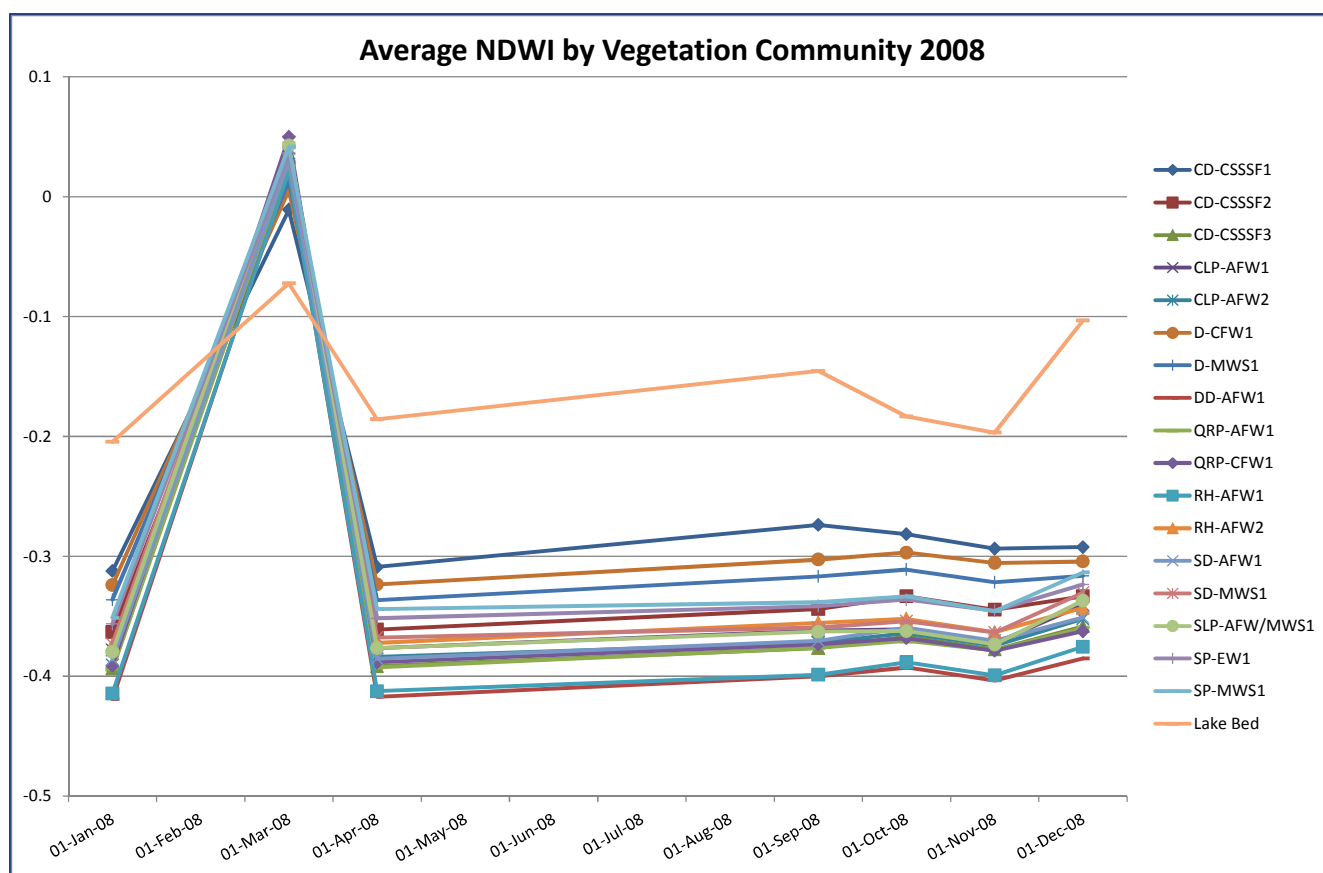


Figure 10 Average NDWI of each vegetation type for 2008

Evapotranspiration (ET)

Cumulative ET

Estimates of actual evapotranspiration (ET) were calculated for the study area using satellite imagery from the CMRSET data set. Cumulative ET was plotted for each year (2006 and 2008) to assess how ET differed over the late wet to late dry season for each of the vegetation types. According to O'Grady et al. (2011) GDE-related vegetation classes are likely to have greater ET losses over the dry period than non-GDE-related classes. The results presented in [Figure 11](#) (2006) and [Figure 12](#) (2008) showed that the highest estimated losses to evaporation are associated with the Lake Bed CD-CSSSF1, which is expected due to surface water and moist soil associated with this unit for a much greater proportion of the dry season than the terrestrial vegetation units. None of the vegetation types displayed consistently high ET rates for either 2006 and 2008. This may indicate that they are not using groundwater and hence are not GDE.

It is also important to note that in 2006 none of the ET values exceed cumulative rainfall indicating that groundwater is not being utilised by these vegetation units. To determine conclusively that this is the case further botanic interpretation should be applied to the study area. It is important to note that the rainfall values used in these figures are for Laverton, which is approximately 180km South West of Lake Wells. Rainfall can be highly variable across this area, so this data may not be fully representative of the project area and any conclusions drawn from this data should be done so with this caveat in mind.

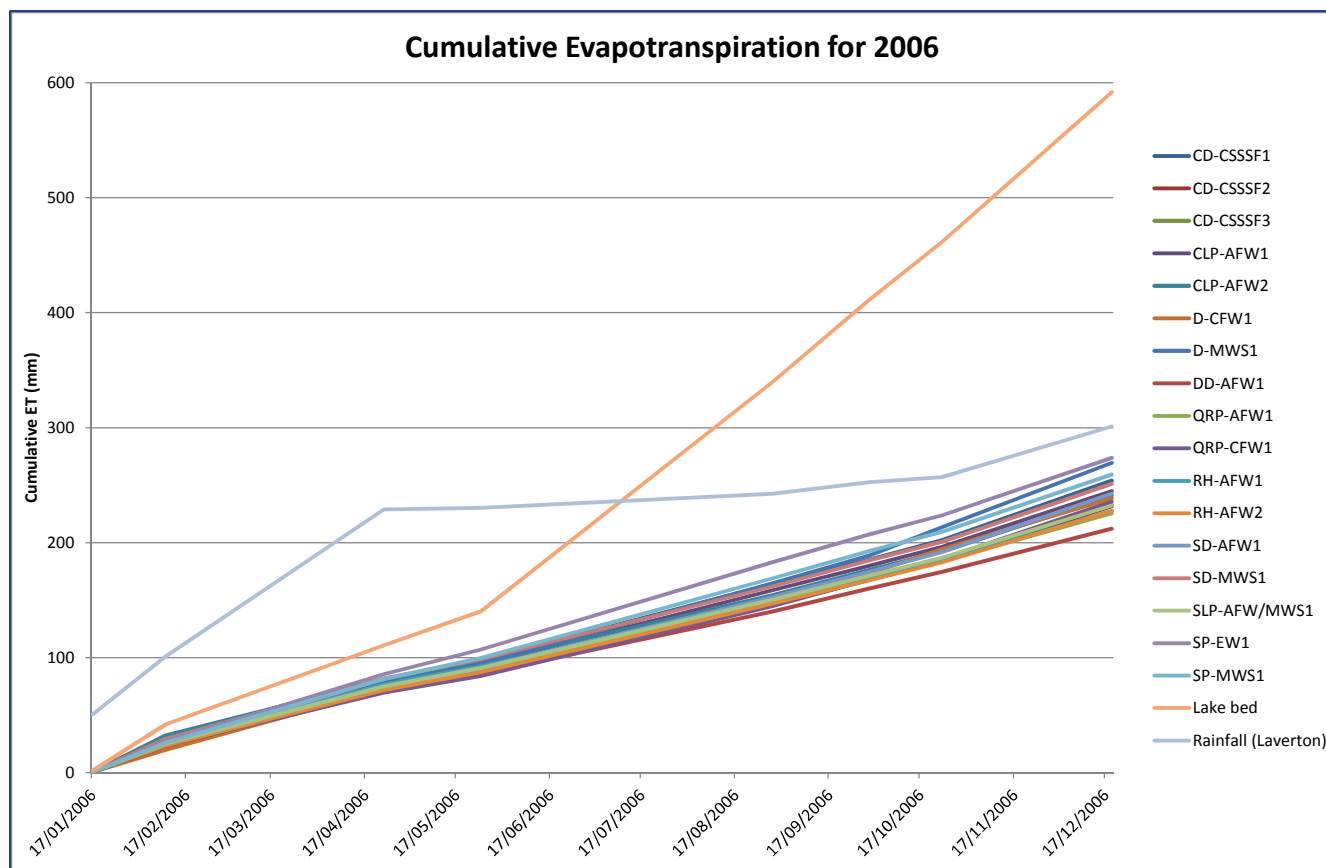


Figure 11 Cumulative ET for each vegetation unit type for 2006

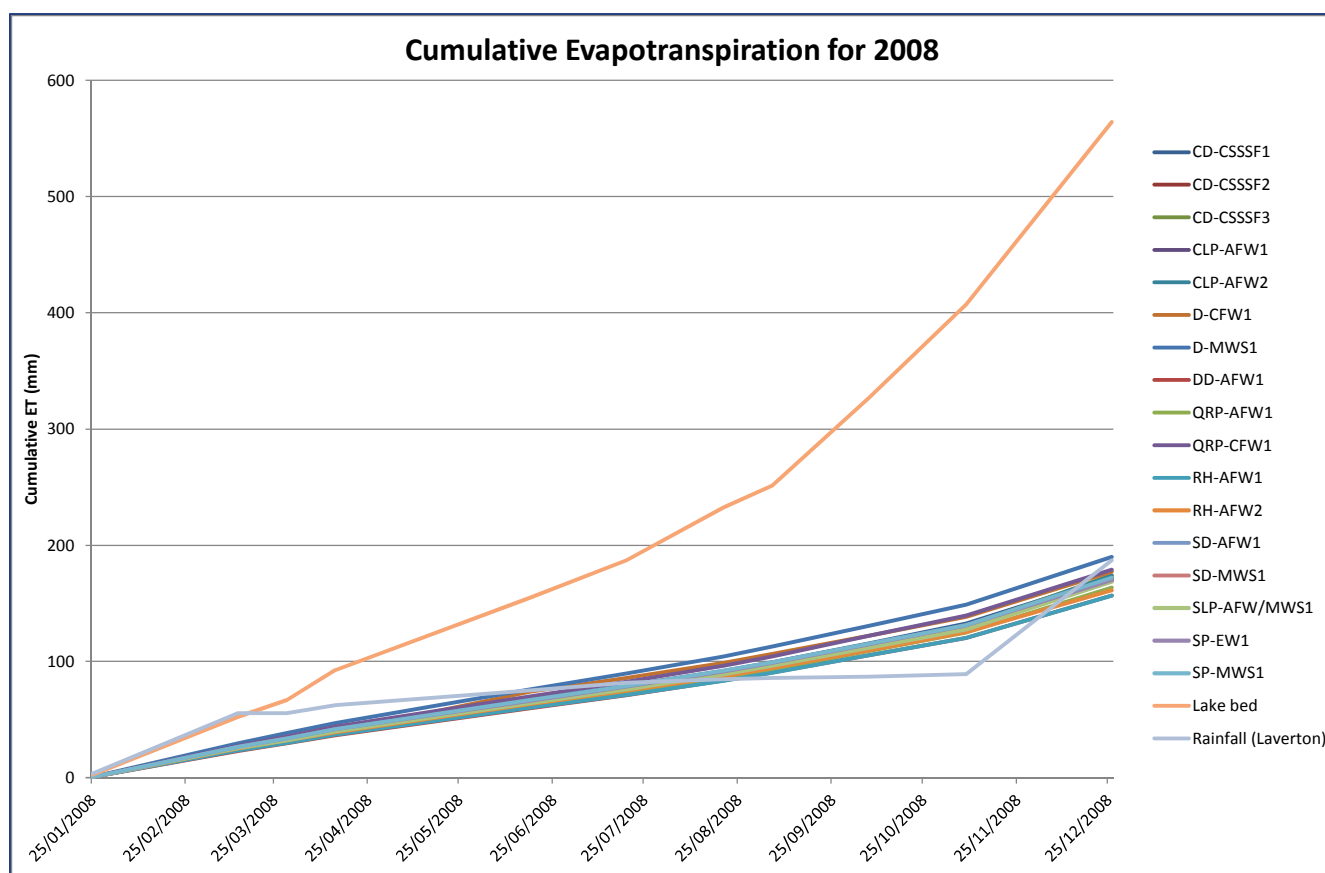


Figure 12 Cumulative ET for each vegetation unit type for 2008

Groundwater Evapotranspiration

The results of the groundwater evapotranspiration assessment are shown in [Figure 13](#). An extension of this method was devised recently by Doody et al. (2017) in which the probability of a vegetation unit using groundwater during dry seasons was calculated, derived from a ratio of ET to rainfall and referred to as the potential inflow dependent ecosystem (pIDE) index. [Table 5](#) shows the ratio and probability of inflow dependence of each vegetation unit (with the exception of the salt lake unit – since this is strictly not a vegetation unit). These results indicate that there is low likelihood of pronounced groundwater dependency in the vegetation units listed in the table.

It is important to note that there are limitations to this method, recent works by van Dijk et al (2015) have shown that the CMRSET method can be unreliable for vegetation units adjacent to salt lake systems and is known to overestimate evapotranspiration in this case.

In addition, the CMRSET data has limited spatial resolution and only allows for a minimum pixel size of 250m². If the size of the vegetation unit is smaller than the pixel size, or is extremely patchy, then values can be skewed based on the values of the other units that may be present. This is of particular relevance for vegetation units next to the salt lake system and due to their 'wet' nature it is highly likely that ET values will be overestimated due to lake bed components being inadvertently included in the ET calculations for that pixel. Consequently caution needs to be taken when making assumptions that a particular vegetation unit is groundwater dependant based solely on estimated ET. To determine conclusively that a unit is a GDE further botanic interpretation should be applied to the study area.

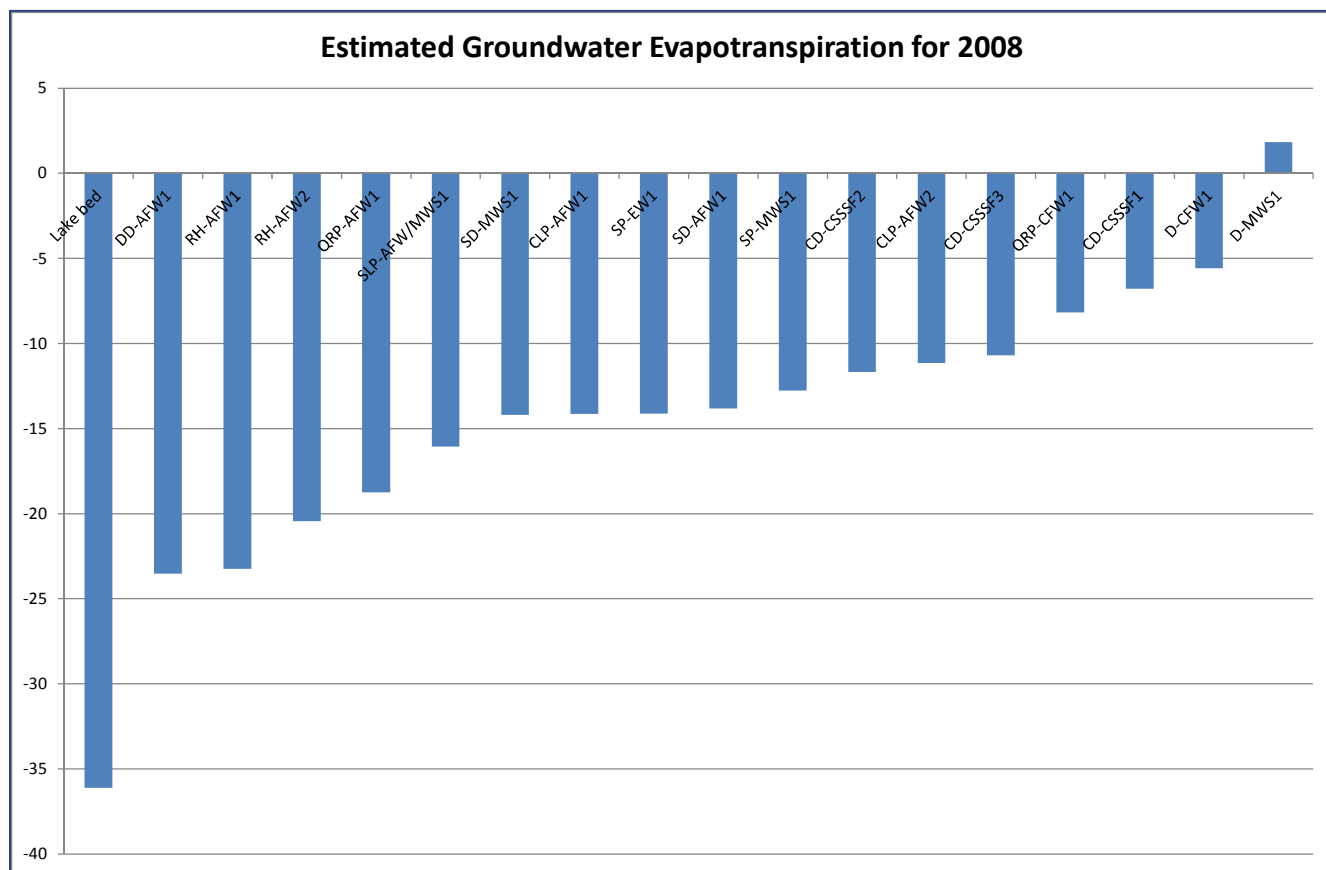


Figure 13 Estimated groundwater evapotranspiration for each vegetation unit (2008)

Table 5 The probability of inflow dependence for each vegetation unit for 2008

Floristic Community	ET/Rainfall Ratio	pIDE (%)
RH-AFW1	0.84	10%
DD-AFW1	0.84	10%
RH-AFW2	0.86	15%
QRP-AFW1	0.87	15%
SLP-AFW/MWS1	0.90	20%
CD-CSSSF2	0.90	20%
SD-AFW1	0.90	20%
SD-MWS1	0.90	20%
SP-EW1	0.91	20%
CLP-AFW1	0.91	20%
SP-MWS1	0.92	20%
CLP-AFW2	0.92	20%
CD-CSSSF1	0.92	20%
CD-CSSSF3	0.93	25%
D-CFW1	0.95	30%
QRP-CFW1	0.95	30%
D-MWS1	1.01	40%

Summary and Conclusions

Spectral data were analysed for two years at Lake Wells.

-) No vegetation unit showed consistently high and unvarying NDVI and NDWI indices (the spectral signature typically associated with groundwater dependent vegetation).
-) Two vegetation units (CD-CSSSF1 and D-CFW1) showed low, but relatively constant NDVI values (for both years) and moderate, but variable, NDWI values (for 2008) (but with lower and less variable wetness than the playa surface). Typically, this signature would indicate areas of sparse vegetation or bare soil.
-) There are methodological issues that limit the application of ET estimation on salt lakes. These limitations constrained the use of ET methods in estimating the likelihood of groundwater dependence of vegetation on islands or in close proximity to the playa.

- J) For vegetation units not closely associated with the salt lake, estimated evapotranspiration and groundwater evaporation amounts did not exceed rainfall, further supporting the conclusion that these units are unlikely to be strongly groundwater dependent.

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